Herbicide acetochlor inhibits adrenergic receptor-mediated calcium oscillations in rat hepatocytes

LIU Fangyuan¹⁾ ZHANG Binbin^{1, 2)†} CUI Zongjie^{1)†}

(1) Institute of Cell Biology, Beijing Normal University, 100875, Beijing, China;

2) Center for Biotechnology Experimental Teaching, School of Pharmacy, Binzhou Medical University, 264003, Yantai, Shandong, China)

Abstract Herbicide acetochlor after application enters both surface and underground waters, so that acetochlor concentrations in rivers and streams could be sufficiently high (nmol \cdot L⁻¹) to exert long-term effects in aquatic animals. Occupational exposure may pose greater risks due to skin and inhalation exposures, resulting in much higher transient blood acetochlor concentrations (μ mol · L⁻¹). Previous works have identified liver as the most susceptible toxicological target for acetochlor, and cytosolic calcium is known to play a vital role in multiple hepatocyte functions. Therefore the present work aims to examine acetochlor effect on calcium oscillations induced by adrenergic receptor activation in freshly isolated rat hepatocytes. It was found that acetochlor at lower concentrations $(1, 10 \, \mu \text{mol} \cdot \text{L}^{-1})$ added simultaneously with phenylepinephrine (PE) had no effect on PE-induced calcium oscillations, but inhibited reversibly calcium oscillations at higher doses (50, 100, 200 µmol · L⁻¹) of acetochlor in selected hepatocytes. Brief prior acetochlor exposure (1, 10, 100 µmol·L⁻¹), however, had no apparent effect on either basal calcium or subsequent PE-induced calcium oscillations. Immunocytochemistry identified non-uniform α1 adrenergic receptor expression among isolated hepatocytes. Immunohistochemistry of liver slices revealed a distinct pattern of α1 adrenergic receptor density in liver lobules: a gradual low-to-high receptor density gradient from pericentral to periportal regions. This receptor density gradient is likely related to acetochlor inhibition in selected hepatocytes. Taken together it is concluded that acetochlor at blood concentrations attainable during acute skin exposure or acetochlor poisoning may exert transient inhibitory effects on α1 receptor signaling in the liver.

Keywords acetochlor; adrenergic receptor; calcium oscillations; hepatocytes

DOI: 10.12202/j.0476-0301.2022145

0 Introduction

Pesticides are widely used to prevent loss in agricultural production. A sizable portion of agricultural pesticides used are herbicides, including chloroacetanilide herbicides such as acetochlor^[1]. Annual acetochlor use both in industrialized US and in developing China is in the tens of millions of pounds^[2–4].

Acetochlor has high water solubility, low soil adsorption, therefore is easily transported from farm fields by surface runoff¹⁵⁻⁶¹. Acetochlor concentrations in surface waters have been found to be in the nanomolar range:

0.926 nmol \cdot L^{-1[7]}, 0.63 nmol \cdot L^{-1[8]}, 1.11 nmol \cdot L^{-1[9]}, 2.85 nmol \cdot L^{-1[10]}, 9.3 nmol \cdot L^{-1[11]}, 0.19~2.67 n mol \cdot L^{-1[12]}, 6.15 nmol \cdot L^{-1[13]}, 3.91 nmol \cdot L⁻¹ (1054.90 ng \cdot L^{-1)[14]}. Atmospheric acetochlor has also been detected, values of up to 53.5 ng \cdot m⁻³ (198.15 nmol \cdot L⁻¹) has been found^[15]. Acetochlor concentration in the rainwater therefore could be up to 9.27 nmol \cdot L^{-1[16]}, 16.26 nmol \cdot L^{-1[17]}. Due to its high soil mobility, the presence of acetochlor in soil has been found to be rather transient^[18–20]. Acetochlor soil runoff into surface water and infiltration to underground water, aerosolized acetochlor precipitated in rain water, together form critical routes of

acetochlor dispersion before moving into the ocean[18, 21-22].

Compared with acetochlor in the environment, occupational exposure to acetochlor poses greater harm to human health. Acetachlor metabolites are readily detected in the urine of farmers^[23] and farm children^[24], with seasonal variations^[23, 25]. Urine acetochlor up to 0.26 nmol·L⁻¹ has been detected in pregnant women residing close to fields of acetochlor application^[26]. Prenatal exposure to acetochlor and other herbicides has been correlated to reduced birth weight in human infants^[27-28].

The effect of long-term and chronic exposure to nanomolar acetochlor has been extensively investigated. It has been found that exposure to environmental concentrations of acetochlor increased thyroid hormone receptor expression in aquatic animals^[29–31]. Acetochlor-spiked feed increased estrogen receptor density in the rat^[32–33]. Acetochlor accelerated frog tadpole metamorphosis via increased expression of the thyroid hormone receptor^[34–35].

Long-term low level acetochlor exposure notwithstanding, effect of short-term occupational exposure at higher concentration has been less well studied. Micromolar acetochlor concentrations have been reported in the serum of acutely poisoned patients (18.61 μ mol · L⁻¹)[36], in skin-exposed monkeys $(0.37 \mu \text{mol} \cdot \text{L}^{-1})^{[37]}$. Further, human skin exposure (at 9.27 mmol · L⁻¹) could lead to contact dermatitis^[38]. Therefore the present work aims to examine acetochlor effect on calcium signal in the freshly isolated rat hepatocytes. Since the al adrenergic receptor plays a prominent role in calcium signaling in hepatocytes[39-43], acetochlor effect on adrenergic receptormediated calcium oscillations was examined. It was found that micromolar acetochlor showed reversible inhibition of PE-induced cytoslic calcium oscillations in hepatocytes. Due to the universal role of calcium signaling in multiple cell types, this new finding has potentially important bearings on the occupational health problems of acetochlor.

1 Materials and methods

1.1 Materials Acetochlor [2-chloro-N-(ethoxymethyl)-N-(2-ethyl-6-methylphenyl) acetamide, MWt 269.77], phenylepinephrine (PE), nuclear stain Hoechst 33 342 were from Sigma-Aldrich (St. Louis, MO, USA). Collagenase H, 2-[4-(2-hydroxyethyl) piperazin-1-yl] ethanesulfonic acid (HEPES) were from Calbiochem

(Darmstadt, Germany). Amino acid mixture (MEM×50) was from Gibco BRL (NY, USA). The cell adhesive Cell-Tak was from BD Biosciences (Bedford, MA, USA). Fura-2 AM was bought from AAT Bioquest (Sunnyvale, CA, USA). Rabbit anti- α_1 adrenergic receptor antibody and TRITC-conjugated donkey anti-rabbit secondary antibody were from Abcam (Cambridge, UK). Acetochlor was dissolved in DMSO, before dilution in buffer to final indicated concentrations. Final DMSO concentration remained < 0.1% which had no effect on either basal calcium or PE-induced calcium oscillations.

1.2 Buffers Standard buffer for isolation of hepatocytes had the following composition (in mmol·L⁻¹): NaCl 118, KCl 4.7, NaH₂PO₄ 1.3, CaCl₂ 2, MgCl₂ 1, HEPES 10, D-glucose 6, L-glutamine 2, pyruvate 2.5, MEM amino acid mixture (50X) 2%, bovine serum albumin (BSA) 0.2%, pH was adjusted to 7.4, and oxygenated for 30 min before use. Perfusion buffer was made without glutamine, pyruvate, amino acid mixture or BSA. For liver perfusion and digestion with collagenase H, collagenase H 24 mg was dissolved in a standard buffer of 80 mL. For filtering out of collagenase H after hepatocytes isolation, cells were layered onto standard buffer containing BSA 2% and centrifuged.

1.3 Isolation of rat hepatocytes Hepatocytes were isolated from male Sprague-Dawley rats (200-400 g) bought from Experimental Animal Facilities, Military Medical Academy Beijing, as reported previously^[39-41, 43]. Briefly, rat was killed and perfused via the portal vein first in calcium-free, heparin-containing (10 IU ⋅ mL⁻¹) buffer (warmed to 37 °C), at a rate of 32 mL ⋅ min⁻¹, then in buffer containing collagenase H (0.29 g ⋅ L⁻¹) at a rate of 20 mL ⋅ min⁻¹. The perfused caudal lobe was excised and digested for a further 2 min in a shaking water-bath (37 °C, 120 r ⋅ min⁻¹). Released hepatocytes were filtered, washed and re-suspended, and judged as > 95% viable by trypan blue exclusion. Only healthy cells were used for experiments as verified by robust and *persistent* calcium oscillations after PE stimulation.

1.4 Measurement of cytosolic calcium concentration in rat hepatocytes Hepatocyte cytosolic calcium concentration was measured as reported previously^[40-41, 43-46]. Briefly, isolated rat hepatocytes were incubated with Fura-2 AM (final concentration 10 μ mol · L⁻¹) for 30 min in a shaking water-bath (37 °C, 50 r · min⁻¹), attached to the

bottom cover-slip (treated with Cell-Tak) of a Sykes-Moore perfusion chamber, and perfused at a rate of $1 \text{ mL} \cdot \text{min}^{-1}$. Cytosolic calcium was measured in a Photon Technology International (PTI, Birmingham, NJ, USA) calcium measurement system. The excitation monochromater was DeltaRam V, with alternate excitation at 340/380 nm. Emitted fluorescence was passed through emitter (510 ± 20) nm and detected with a Hamamatsu PMT attached to an Olympus IX70 inverted fluorescence microscope. Fluorescence ratios recorded (1 ratio per min) were plotted with SigmaPlot, as representative of cytosolic calcium changes as reported before [43, 45, 47–49].

To detect any effect of acetochlor on the perifused rat hepatocytes, hepacocytes were first stimulated with the $\alpha 1$ adrenergic agonist PE, then acetochlor at different concentrations were added for 10 min, followed by a second dose of PE. Alternatively, hepatocytes were perfused throughout, acetochlor was then added for 15 min on top of PE stimulation. For statistical analysis, calcium spikes before, during, and after acetochor was counted and plotted as mean \pm SEM (number of spikes in 15 min). Statistical significance was noted when P < 0.05. All experiments were repeated at least 3 times with hepatocytes from \geq 3 different rats killed on different days each.

1.5 Immunocytochemical and immunohistochemical demonstration of graded all receptor density in isolated hepatocytes and in liver lobule Freshly isolated rat hepatocytes were fixed in paraformaldehyde 4% for 10 min, permeabilized in 0.2% Triton X-100 for 15 min, nonspecific binding blocked in 3% BSA for 60 min, then incubated with rabbit anti-α1 adrenergic receptor antibody (1:100) at 4 °C for 12-14 h, TRITC-conjugated donkey anti-rabbit secondary antibody (1:100) at room temperature for 1 h. Controls were done: without primary and secondary antibodies, with primary but without secondary antibodies, without primary but with secondary or with non-specific IgG and secondary antibodies, antibody. All controls showed negative signal. Cell nucleus was visualized with Hoechst 33342. All images were taken in a laser scanning confocal microscope (Zeiss 510 Meta) under objective×63, with TRITC at λ_{ex} 543 nm/ λ_{em} 590 nm, Hoechst 33 342 at λ_{ex} 405 nm/ λ_{em} 420 nm. The α1 adrenergic receptor fluorescence density in hepatocytes (#1-#10) was quantified with Aim-ImageBrowser, and plotted sequentially with SigmaPlot.

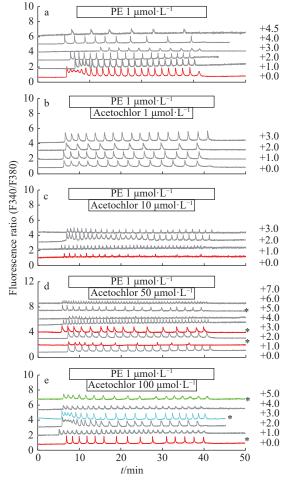
For immunohistochemistry, rat liver was excised and fixed in paraformaldehyde 4% (4 °C, 12 h), dehydrated in 30% saccharose (4 $^{\circ}$ C, 12 h), and slices (10 μ m thick) frozen cut. The slices were first incubated in 3% BSA for 60 min to block nonspecific binding, then incubated with rabbit anti-α1 adrenergic receptor antibody (1:100, 4 °C 12-14 h), FITC-conjugated goat anti-rabbit secondary antibody (1:100) at room temperature for 1 h. Controls were done: without primary and secondary antibodies, with primary but without secondary antibodies, without primary but with secondary antibodies, or with nonspecific IgG and secondary antibody. All controls showed negative signal. Cell nucleus was stained with Hoechst 33342. Images were taken in a laser scanning confocal microscope (Zeiss 510 Meta) under objective \times 20 or \times 10, with FITC at $\lambda_{\rm ex}$ 488 nm/ $\lambda_{\rm em}$ 495 nm; Hoechst 33342 at λ_{ex} 405 nm/ λ_{em} 420 nm. The $\alpha 1$ receptor fluorescence intensity along lines in neighboring pericentral and periportal regions and along hepatocyte stacks in a liver lobule was quantified with AimImageBrowser and plotted with SigmaPlot.

2 Results

2.1 Acetochlor inhibition of phenylepinephrine (PE)-induced calcium oscillations in selected hepatocytes

Phenylepinephrine (PE) at low concentrations induced cytosolic calcium oscillations in the freshly isolated rat hepatocytes, as reported before [39-41]. PE 1 μ mol · L⁻¹ when added to perifused rat hepatocyte triggered regular calcium oscillations, note the intrinsic variation in oscillatory frequency in different hepatocytes (Fig. 1-a). Addition of acetochlor at 1 µmol · L⁻¹ had no apparent effect on PE-induced calcium oscillation (Fig. 1-b), neither did acetochlor at 10 μmol·L⁻¹ (Fig. 1-c). With acetochlor 50 µmol·L⁻¹ added, the frequency of PEinduced calcium oscillations was minimally diminished in 3 out of 8 experiments, with slight recovery after wash-out (tracings marked with *, Fig. 1-d). This inhibition was more marked with acetochlor 100 µmol · L⁻¹. In 3 out of 6 experiments acetochlor 100 µmol · L⁻¹ reduced calcium oscillatory frequency with gradual recovery after wash-out of acetochlor. The oscillatory frequency of PE-induced calcium oscillations were (0.63 ± 0.09) , (0.36 ± 0.02) , (0.60 ± 0.06) spikes/min before, during and after acetochlor 100 µmol · L⁻¹ exposure, i.e., with significant inhibition in these hepatocytes (tracings marked with *, Fig. 1-e, P < 0.05, N = 3). Some inhibitory effect was also found in 3 other experiments, but the inhibition was less pronounced (Fig. 1-e, N = 3). From these experiments it is clear that acetochlor at low concentrations (1 and 10 µmol · L⁻¹) had no apparent effect on PE-induced cytoslic calcium oscillations, but at higher concentrations (50, 100 µmol · L⁻¹) acetochlor reduced the oscillation frequency, but only in selected heaptocytes.

To confirm that acetochlor inhibition shown above



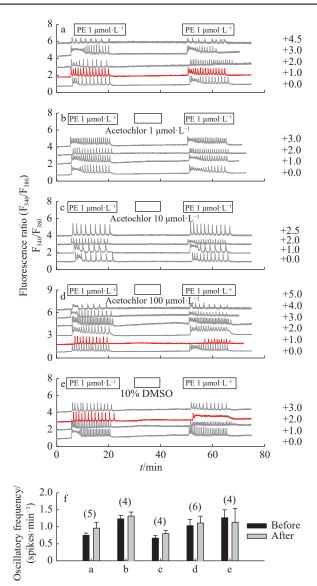
Fura-2 AM-loaded rat hepatocytes were perifused, phenylepinephrine (PE, 1 μ mol \cdot L⁻¹), acetochlor at indicated concentrations were added as marked by the horizontal bars. (a) Control experiment with continuous PE (1 μ mol \cdot L⁻¹) stimulation. Acetochlor 1 μ mol \cdot L⁻¹ (b), 10 μ mol \cdot L⁻¹ (c), 50 μ mol \cdot L⁻¹ (d), 100 μ mol \cdot L⁻¹ (e) was added during PE (1 μ mol \cdot L⁻¹) stimulation. Calcium tracings for each experiment are drawn together after adding the indicated values (+0.0, 1.0, 2.0, 3.0, 4.0, 4.5, 5.0, 6.0, 7.0) to measured F₃₄₀/F₃₈₀ ratios. Asterisks (*) in panels (d) and (e) indicate observed acetochlor inhibition.

Fig. 1 Inhibition by acetochlor of α1 adrenergic receptormediated calcium oscillations in selected hepatocytes

(Fig. 1-d, e) was not due to solvent effect, vehicle DMSO 0.1% (the highest concentration used when acetochlor was at 100 μ mol \cdot L⁻¹) was added, which showed no effect on PE-induced calcium oscillations (not shown).

2.2 Brief acetochlor exposure showed no lasting inhibition on subsequent PE-induced calcium oscillations in the freshly isolated rat hepatocytes Repeated stimulations of micromolar PE induced reproducible calcium oscillations in perifused rat heaptocytes. The two sequential PE challenges resulted in identical or highly similar pattern of calcium oscillations (Fig. 2-a). In between the two PE stimulations, acetochlor exposure at 1-100 μmol · L⁻¹ had no effect on either basal calcium (Fig. 2-b-d), or on calcium oscillations induced by a second dose of PE (Fig. 2-b-d). For comparison, DMSO at 10% did not change basal calcium concentration (Fig. 2-e), neither did it inhibit significantly the second PE stimulation in 2 experiments, but either resulted in an elevated plateau or diminished oscillations in 2 other experiments (Fig. 2-e). Therefore acetochlor at 1-100 μmol · L⁻¹ had no lasting effect on cytosolic calcium in hepatocytes but DMSO 10% did show minor modification of subsequent calcium oscillations (Fig. 2-f).

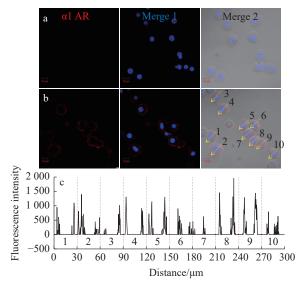
2.3 Graded al adrenergic receptor density among freshly isolated rat hepatocytes The calcium experiments above indicated that acetochlor tended to inhibit PEinduced calcium oscillations in hepatocytes which showed lower oscillatory frequency after PE stimulation at 1 μmol · L⁻¹ (Fig. 1). Lower frequency of calcium oscillation could either be due to low PE concentration, or due to low density of al adrenergic receptor. Since PE concentration used was the same, we decided to examine possible variations in all adrenergic receptor density in the freshly isolated rat hepatocytes by immunocytochemsitry. Fig. 3 reveals varied α1 adrenergic receptor density in different hepatocytes. Control experiments indicated that without anti-α1 adrenergic receptor antibody, no staining was found (Fig. 3-a). The $\alpha 1$ adrenergic receptor was mainly localized on the plasma membrane in the isolated hepatocyets with the clearly defined nuclear and cytosolic regions without staining (Fig. 3-b). The receptor density was markedly different among isolated hepatocytes (Fig. 3-b). Of those hepatocytes shown in Fig. 3-b, the receptor density was the highest in cell number 8, and lowest in cell number 7, with the rest in between: 8 > 9 > 2 > 5 > 4 >



Fura-2 AM-loaded rat hepatocytes were perifused, PE (1 μ mol \cdot L⁻¹), acetochlor (1, 10, 100 μ mol \cdot L⁻¹), or 10% DMSO were added as indicated by the horizontal bars. (a) Control: 2 sequential PE 1 μ mol \cdot L⁻¹ stimulations. Acetochlor 1 μ mol \cdot L⁻¹ (b), 10 μ mol \cdot L⁻¹ (c), 100 μ mol \cdot L⁻¹ (d), 10% DMSO (e) were added in between 2 PE 1 μ mol \cdot L⁻¹ stimulations. Calcium tracings for each experiment are drawn together after adding the indicated values (+0.0, 1.0, 2.0, 2.5, 3.0, 4.0, 4.5, 5.0) to measured F₃₄₀/F₃₈₀ ratios. (f) The frequencies of calcium oscillation in panels (a-e) were analyzed. In each case (a, e) frequency of first PE stimulation ("Before") was not significantly different from that of second PE stimulation ("After"). No statistical significance was detected.

Fig. 2 Acetochlor or DMSO had no effect on sequential stimulation of α1 adrenergic receptor by PE in the freshly isolated rat hepatocytes

1 > 3 > 6 > 10 > 7 (Fig. 3-c). Note that in the fluorescent density plot, each hepatocyte is indicated by 2 major cluster peaks, representing the opposing plasma



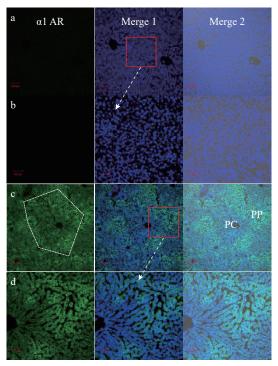
Each panel is composed of micrographs of $\alpha 1$ adrenergic receptor fluorescence (red), nuclear fluorescence (blue), and bright-field merged images. Freshly isolated rat hepatocytes were fixed, incubated without (a) or with (b) rabbit anti- $\alpha 1$ adrenergic receptor, then with TRITC-conjugated donkey anti-rabbit secondary antibody (red), and double stained with nuclear stain Hoechst 33 342 (blue). The $\alpha 1$ receptor fluorescence intensity along the dotted lines for each hepatocyte was quantified with software AimImageBrowser, and plotted sequentially from hepatocyte # 1 to # 10 with SigmaPlot (c). Confocal images were obtained in a Zeiss laser scanning microscope (Zeiss LSM 510 META) under objective 63 x / 1.40 oil. Scale in image is 10 μ m. Images shown are representative of 3 identical experiments (N = 3).

Fig. 3 Graded α1 adrenergic receptor density in freshly isolated rat hepatocytes

membrane patches along the scan line (Fig. 3-c).

2.4 A clear adrenergic all receptor density pattern in the liver lobule To correlate receptor density with in situ localization of hepatocyte, immunohistochemistry was performed in thinly-cut liver slices. In such slices, lobule was found to be composed of hepatocytes arranged radially, surrounding neighboring portal triad or central vein. The α1 adrenergic receptor density among neighboring hepatocytes in the liver lobule in frozen-cut liver slices showed typical graded receptor density (Fig. 4). Control experiments without added primary antibody but with secondary antibody did not show any staining for the α1 adrenergic receptor (Fig. 4-a, 4-b). The lower panel micrographs (Fig. 4-b, 4-d) are the expanded portions from the upper panels (Fig. 4-a, 4-c). The α 1 adrenergic receptor density varied depending on the hepatocyte location in the lobule (Fig. 4-c, 4-d). The highest density was found in the periportal (PP) region (marked by the dashed lines in Fig. 4-c), with receptor density gradually decreasing towards the pericentral (PC) region (Fig. 4-c, 4-d).

Such graded α1 adrenergic receptor density was analyzed and presented quantitatively in Fig. 5-A and 5-B. The micrograph in Fig. 5-Aa is taken from the rectangular part in Fig. 4-c. Tracings 1-6 in Fig. 5-Ab show that in each and every direction (lines 1-6) starting from the central vein (pericentral, PC), α1 adrenergic receptor density increased gradually in a pericentral-to-periportal (PC-to-PP) direction, reaching a maximum at the periportal (PP) triad. A line drawn between 2 central veins (Fig. 5-Aa, line 7) showed an overall "mountain peak"



Each panel is composed of α1 adrenergic receptor fluorescence, merged α1 adrenergic receptor/nuclear fluorescence (merge 1), and merged α1 adrenergic receptor/nuclear fluorescence/bright field (merge 2) images. Frozen rat liver slices (10 μm thick) were cut, incubated without (a, b) or with (c, d) rabbit anti-α1 adrenergic receptor antibody, then with FITC-conjugated goat anti-rabbit secondary antibody (green), and double stained with Hoechst 33342 (blue). Images were taken in a laser scanning confocal microscope (Zeiss LSM 510 META) under objectives (in magnification/numerical aperture) 10 x/1.40、20 x/1.40, scales shown are 50 or 100 μm. Panels b and d were taken from Panels a and c respectively as outlined by the red squares. Images shown are representative of 4 identical experiments (N = 4).

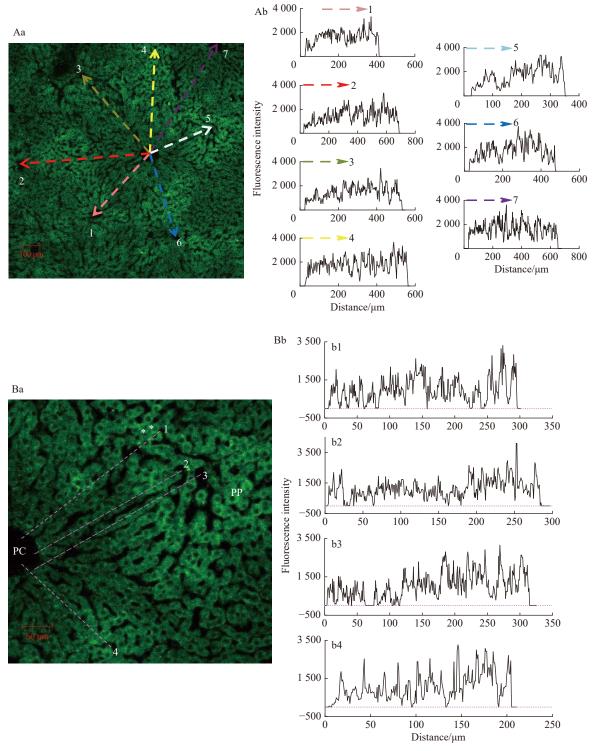
Fig. 4 Graded density distribution of α_1 adrenergic receptor in rat liver slice

structure (Fig. 5-Ab, tracing 7).

Fig. 5-Ba (taken from Fig. 4-d) shows clear stacks of hepatocytes. Line scans (lines 1-4) along such hepatocyte stacks indicate that the $\alpha 1$ adrenergic receptor density (fluorescence intensity) was the highest in hepatocyte at the periportal end of the stack and lowest in hepatocyte at the pericentral end of the stack (Fig. 5-Bb, tracings 1-4). Distinct single hepatocyte profiles could be clearly recognized in such line scan tracings in that each hepatocyte is composed of 2 peaks representing that the scan line crossing the opposing plasma membranes; see the asterisks in both micrograph and tracing b1, for example. Although the reason for such graded $\alpha 1$ adrenergic receptor density is not clear, it is obvious that nutrient/hormone gradients are likely to play a major role.

3 Discussion

It has been found in the present work that when acetochlor was added to PE-stimulated rat hepatocytes, low acetochlor concentrations (1, 10 µmol · L⁻¹) had no effect on PE-induced calcium oscillations, but higher concentrations (50, 100 µmol·L⁻¹) showed reversible inhibition in selected hepatocytes which showed lower oscillatory frequency. Acetochlor (1-100 µmol · L⁻¹) had no effect on basal cytosolic calcium, nor did it have any effect on calcium oscillations induced by a second dose of PE. Acetochlor inhibition was reversible and the inhibition was seen only in selected hepatocytes which showed lower oscillatory frequencies. The selective inhibition was interesting and could be due to lower al adrenergic receptor density levels in hepatocytes from the pericentral region of the liver lobule. The increasing α1 adrenergic receptor density from pericentral-to-periportal portion of the liver lobule denotes that hepatocytes close to the central vein showed the lowest receptor density whereas hepatocytes close to the portal triad (hepatic artery, portal vein, bile duct) showed the highest receptor density. Lower receptor density results in lower frequency of calcium oscillation after receptor activation, therefore is more susceptible to acetochlor inhibition. These data together indicate that acetochlor had reversible inhibitory effect on hepatocyte al adrenergic receptor, but only in selected hepatocytes probably corresponding to the central vein region in the liver lobule. Future live liver slice or in situ calcium imaging likely will explicitly confirm such



A. Fluorescence quantification of α1 adrenergic receptor distribution in rat liver lobules. The main image is from Fig. 4-c, fluorescence intensity was scanned as indicated by the colored arrows, from pericentral (PC) to periportal (PP) regions. Purple arrow transcends two neighboring pericentral regions. Fluorescence intensity was quantified with AimImageBrowser, and plotted with SigmaPlot. Data shown are typical of N identical experiments (N = 6). B. Fluorescence quantification of α1 adrenergic receptor distribution in hepatocyte stacks of the rat liver lobule. The main image is from Fig. 4-d, fluorescence intensity was scanned as indicated by the numbered dashed lines, in the pericentral (PC)-to-periportal (PP) direction. Fluorescence intensity was quantified with AimImageBrowser, and plotted with SigmaPlot. Note the 2 white asterisks at the periportal end of the dashed line 1 in the micrograph and the last 3 cluster peaks between 250 - 300 μm in tracing b1 denote the location of 2 neighboring periportal hepatocytes. The thin dashed pink lines in tracings b1-b4 indicate blank area with zero fluorescence.

Fig. 5 A pericentral-to-periportal gradient of α1 adrenergic receptor density in rat liver lobule

findings.

The α 1 adrenoceptor is a major receptor involved in calcium signaling^[40–42], and plays a pivotal role in hepatocyte metabolism, such as hepatic glucose production^[42, 50–51] and liver regeneration^[52]. Therefore acetochlor interference of α 1 adrenergic receptor calcium signaling bears special significance, especially when *in vivo* or *in situ* acetochlor concentration reached high levels (18.61 μ mol · L⁻¹)^[36].

Although toxic effect of acute high acetochlor concentration has not been systematically examined, chronic effect of nanomolar acetochlor has been well investigated. Nanomolar acetochlor (10 nmol \cdot L⁻¹) has been found to influence tadpole metamorphosis by increasing thyroid hormone receptor expression and to accelerate T3-induced metamorphosis in *Xenopus laevis* tadpoles^[34–35].

Long-term (21 d) acetochlor exposure (0.07, 0.74, 7.41 nmol · L⁻¹) has been found to enhance liver expression of the thyroid hormone receptor α in adult rare minnow (Gobiocypris rarus), acetochlor at lower concentration (0.07 nmol · L⁻¹) enhanced but at higher concentrations (0.74, 7.41 nmol · L⁻¹) diminished brain thyroid hormone receptor α expression^[31]. Acetochlor feeding (5 mL·kg⁻¹ in sunflower seed oil, 20% LD₅₀ for 6 d) increased estrogen-binding capacity in uniparous female rat^[32]. Acetochlor feeding (7.58, 15.36 mg \cdot kg⁻¹ \cdot d⁻¹, 6 d) also increased uterine nuclear estrogen receptor density[33]. Significant changes in gene (betaine homocysteine methyltransferase, chemotaxin LECT2, and apolipoprotein E1) expression in liver have been detected after 62 d exposure to a mixture of herbicides (total concentration 4 $\mu g \cdot L^{-1}$) including acetochlor (0.5 $\mu g \cdot L^{-1}$ or 1.85 nmol \cdot L⁻¹) in the European flounder *Platichthys flesus*^[53].

Early toxicological work at mega doses of acetochlor (peri-maximum tolerated dose, 1000 mg · kg⁻¹ or 1000 ppm in diet) found that acetochlor was not genotoxic but was epigenetically carcinogenic, and identified three susceptible toxicological targets: human lymphocytes, rat (but not mouse) nasal epithelium, and rat liver^[54]. The active group in the acetochlor molecule was found to be the chloroacetyl (— COCH₂Cl) substituent, which targets sulfhydryl (—SH) groups in glutathione (GSH) and protamine^[54–55]. The targeted action of acetochlor was highlighted further by the fact that rat sperms were remarkably resistant to acetochlor toxicity; even when rats

were treated with maximum or supra-maximum tolerated doses, no mutagenic or even histopathological effects were observed^[55].

Acetochlor administration at maximum tolerated dose over extended time period (2 year feeding) induced nasal adenoma in the rat^[54, 56]. Nasal epithelium accumulation of an acetochlor metabolite sulfoxide quinine-imine after oral doses, and its covalent-bonding to Cys free sulfhydryl groups (Protein-Cys-SH) were noted to be important for adenoma induction^[56]. Blood plasma analysis indicated that hepatocyte metabolite acetochlor sulfoxide was the major circulating metabolite available to nasal epithelium, where it was further metabolized to the reactive sulfoxide quinine-imine^[56].

Supra-maximum tolerated induced dose also occasionally thyroid tumors in the rat, which were associated with increased levels of thyroid stimulating hormone (TSH) and thyroid weight, thyroid hormone T3^[54, 57]. This could be due to the antithyroid effect of acetochlor. It was found that acetochlor enhanced uridine diphosphate glucuronosyltransferase (UDPGT) activity in liver therefore increased hepatic thyroid hormone metabolism and excretion; lowered thyroid hormone level led to persistent stimulation of thyroid follicular cell therefore thyroid hyperplasia and carcinogenesis^[57].

Possible herbicide effect on calcium signaling has not been investigated in detail before. Herbicide 2, 4-dichlorophenoxyacetic acid has been reported to have no effect on cytosolic calcium in the isolated CD4+ T lymphocytes, but when lymphocytes were previously activated with soluble anti-CD3 antibody, 2, 4-dichlorophenoxyacetic acid (\geq 100 µmol·L⁻¹) stimulated irreversible and persistent (1 to 24 h) calcium increase^[58]. Such calcium increases may be mediated by 2, 4-dichlorophenoxyacetic acid-stimulated production of hydrogen peroxide^[58].

The acute effect of acetochlor has been examined in neurons. Acetochlor (100, 200, 400, 800 μ mol · L⁻¹) exposure over extended time periods (24, 50 h) was found to decrease the amplitude of field stimulation-elicited compound action potential in sciatic nerve of the frog *Rana ridibunda*, both exposure time- and concentration-dependently^[59]. At physiological pH the IC₅₀ was found to be 220 μ mol · L⁻¹, the dose-response curve being so steep

that acetochlor at 100 μ mol · L⁻¹ was without any effect^[59]. At pH 3.3, the sciatic nerve was severely compromised, IC₅₀ was found to be 770 μ mol · L⁻¹ but again acetochlor at 600 μmol · L⁻¹ was without any effect^[59]. In shorter time incubations the neurons survived better. The shortest time point examined was 1 h, acetochlor at ≤ 400 micro mol · L⁻¹ had no effect, but at 800 μmol · L⁻¹ only 65% nerve vitality remained^[59]. In this interesting work, no molecular mechanisms (ionic channel involvement or nerve cell plasma membrane permeability changes) were examined. Acetochlor effect on cell vitality has also been examined in cell lines Vero, HepG2, MCF7, cytotoxic IC_{50} was found to be 57, 116, and 231 µmol · L⁻¹ (15.6, 32.1, 62.5 μ g · mL⁻¹) respectively^[60]. It may be noted that LC₅₀ (in 48 h) of acetochlor in invertebrates has been found to be in the mmol · L⁻¹ range (0.47-4.37 $\text{mmol} \cdot L^{-1})^{[61-62]}$.

Acetochlor (10, 100, 1000 μ mol·L⁻¹) stimulated the multidrug resistance (MDR) type 1 (MDR1) (an ATP-binding cassette or ABC transporter) ATPase activity with an EC₅₀ of 25 μ mol·L⁻¹, inhibited calcein efflux with an IC₅₀ > 100 μ mol·L⁻¹, but had little effect on other ABC transporters (MRP1, MRP2, BCRP)^[1]. Acetochlor stimulation of MDR1 activity sensitized MDR1-K562 cells to cytostatic paclitaxel^[1].

Prolonged acetochlor exposure (for many h) had different effects. Acetochlor up to 10 μmol · L⁻¹ (48 h) had no effect on the viability of human breast cancer cell lines MCF-7, MCF-10A, MDA-MB-231^[63]. But acetochlorat1mmol · L⁻¹inducedsignificant(6%) humanerythrocyte hemolysis (in 24 h)^[64]. Acetochlor exposure (25-800 μmol · L⁻¹) to human lung carcinoma cell A549 resulted in reactive oxygen species generation (in 2-3 h) and apoptosis (in 24-120 h)^[65]. Acetochlor was found to have little cytotoxicity in fibroblasts (AS52) at 10 μmol · L⁻¹, although at 100 μmol · L⁻¹ significant toxicity was detected; acetochlor at both 4 and 40 μmol · L⁻¹ increased the proportion of cells at G1 stage of the cell cycle, with a corresponding decrease in G2 phase^[66].

As mentioned above the most susceptible cell type in *in vivo* studies in rat has been found to be hepatocytes^[54]. The hepatocyte toxicity of acetochlor (at 100, 200, 400, 800, $1\,600~\mu\text{mol} \cdot L^{-1}$) has been examined by measuring lactate dehydrogenase (LDH) leakage from intact hepatocytes. It was found that acetochlor at $100~\mu\text{mol} \cdot L^{-1}$ had

no effect after incubation with isolated rat hepatocytes for 2 h, cytotoxic LC₅₀ (50% toxicity) was 706 μ mol · L^{-1[2]}. In two lots of cryo-preserved human hepatocytes, acetochlor at 200 or 400 μ mol · L⁻¹ had no effect after 2 h incubation; the cytotoxic LC₅₀ was found to be 587, 1348 μ mol · L⁻¹ respectively^[2]. These data clearly indicated that acetochlor was quite non-toxic towards hepatocytes even at 100 μ mol · L⁻¹ over an extended incubation period (2 h). Although acetochlor (1-100 μ mol · L⁻¹) was non-toxic for freshly isolated rat hepatocytes, it did inhibit PE-induced calcium oscillations reversibly in selected hepatocytes (Fig. 1).

The present work found that not all hepatocytes were susceptible to acetochlor inhibition. Reversible inhibition by acetochlor was seen only in hepatocytes showing more sparse calcium oscillations (Fig. 1), likely due to lower α 1 adrenergic receptor density (Fig. 3, 4). Since hepatocytes with sparse calcium oscillations are likely pericentral (Fig. 4, 5) therefore more distant to the circulating portal blood, in whole organisms pericentral hepatocytes may be less affected than periportal hepatocyte by circulating acetochlor. The fact that all adrenergic receptor is present in the highest density in periportal region (Fig. 4, 5) is consistent with reports that calcium waves always start from the periportal region to spread to pericentral hepatocytes in the liver^[67–68]. To our best knowledge, present work is the first fluorescent immunohistochemical report of a pericentral-to-periportal gradient of α1 adrenergic receptor density (Fig. 4, 5). In the present work it was not possible to correlate hepatocyte receptor density in frozen cut liver slice (10 µm thick) directly with calcium oscillation. In the future it will be interesting to invesin live liver sections, to correlate modulated calcium signaling with receptor density, and possibly also with density of other signaling proteins.

Although brief incubation with acetochlor 100 μ mol · L⁻¹ had no effect on basal calcium concentration in the freshly isolated rat hepatocytes (Fig. 2) and showed rather reversible inhibition of PE-induced calcium oscillations (Fig. 1), acetochlor 100 μ mol · L⁻¹ has been reported to increased cytosolic calcium concentration markedly with extended incubation period of 24 h in the human liver carcinoma cell line HepG2^[69]. Acetochlor 20 mg · L⁻¹ (74.1 μ mol · L⁻¹) was found to induce apoptosis and related signaling in a large majority of HepG2 cells over a

24 h period^[70]. Continued monitoring of cytosolic calcium concentration over extended time period (24 h or longer) might give some details as to the time course of calcium changes, it must be born in mind that acetochlor might not be cleared out of the body very quickly in the exposed patients.

As mentioned above, long-term exposure (2 year feeding) of acetochlor in rat leads to nasal tumor, largely due to persistent presence in nasal epithelium of a sulfoxide intermediate. Hydroxylation of the phenyl ring of the sulfoxide intermediate to quinine-imine was important for hyperplasia induction in nasal epithelium^[56]. This reactive quinine-imine/quinonimine could react covalently with the sulfhydryl groups in protein side chain to induce protein modification^[56, 71]. Metabolite studies in rat liver microsomes found that acetochlor was metabolized sequentially to 2-cholor-N-(2-methyl-6-ethylphenyl) acetamide (CME-PA), 2-methyl-6-ethylaniline (MEA), and dialkylbenzoquinone^[72]. Whether adrenoceptor Cys thiols would play any role during brief incubation with acetochlor in perfused heaptocytes is not known at present. Free thiol oxidative modulation of receptor may play a role in acetochlor inhibition of the al adrenoceptor signaling.

In conclusion, acetochlor at high concentrations inhibits αl adrenoceptor-mediated calcium signaling in selected hepatocytes, which may have implications for occupational exposure and acute poisoning. This acute effect on αl adrenoceptor adds to the well-recognized long-term effect on thyroid hormone receptor and estrogen receptor.

Acknowledgments:

Work supported by the Natural Science Foundation of China (NSFC, 31971170), and by A Visiting Scholarship from Shandong province (to Zhang Binbin), China.

4 References

- [1] OOSTERHUIS B, VUKMAN K, VÁGI E, et al. Specific interactions of chloroacetanilide herbicides with human ABC transporter proteins[J]. Toxicology, 2008, 248(1): 45
- [2] KALE V M, MIRANDA S R, WILBANKS M S, et al. Comparative cytotoxicity of alachlor, acetochlor, and metolachlor herbicides in isolated rat and cryopreserved human hepatocytes[J]. Journal of Biochemical and Molecular Toxicology, 2008, 22(1):41
- [3] 丁丽,付颖,叶非. 酰胺类除草剂的研究和应用进展[J]. 农

- 药科学与管理,2011,32(9):22
- [4] 柳梅,王天斌,沈小德. 我国乙草胺市场现状及技术进展 [J]. 化工技术经济,2005,23(3):14
- [5] VECCHIA A V, GILLIOM R J, SULLIVAN D J, et al. Trends in concentrations and use of agricultural herbicides for Corn Belt rivers, 1996-2006[J]. Environmental Science & Technology, 2009, 43(24): 9096
- [6] 唐易, 杨浩娜, 彭亚军, 等. 乙草胺迁移降解规律研究进展 [J]. 农药, 2021, 60(5): 313
- [7] CAPEL P D, MA L, SCHROYER B R, et al. Analysis and detection of the new corn herbicide acetochlor in river water and rain[J]. Environmental Science & Technology, 1995, 29(6): 1702
- [8] BOYD R A. Herbicides and herbicide degradates in shallow groundwater and the Cedar River near a municipal well field, Cedar Rapids, Iowa[J]. Science of the Total Environment, 2000, 248(2/3); 241
- [9] CLARK G M, GOOLSBY D A. Occurrence and load of selected herbicides and metabolites in the lower Mississippi River[J]. Science of the Total Environment, 2000, 248(2/3):101
- [10] KOLPIN D W, THURMAN E M, LINHART S M. Finding minimal herbicide concentrations in ground water? Try looking for their degradates[J]. Science of the Total Environment, 2000, 248(2/3): 115
- [11] BATTAGLIN W A, FURLONG E T, BURKHARDT M R, et al. Occurrence of sulfonylurea, sulfonamide, imidazolinone, and other herbicides in rivers, reservoirs and ground water in the Midwestern United States, 1998[J]. Science of the Total Environment, 2000, 248(2/3): 123
- [12] SCRIBNER E A, BATTAGLIN W A, GOOLSBY D A, et al. Changes in herbicide concentrations in Midwestern streams in relation to changes in use, 1989-1998[J]. Science of the Total Environment, 2000, 248(2/3); 255
- [13] REBICH R A, COUPE R H, THURMAN E M. Herbicide concentrations in the Mississippi River basin: the importance of chloroacetanilide herbicide degradates[J]. Science of the Total Environment, 2004, 321(1/2/3): 189
- [14] YUZY, JINF, LIHY, et al. Residual levels of acetochlor in source water and drinking water of China's major cities[J]. Environmental Science, 2014, 35(5): 1694
- [15] PECK A M, HORNBUCKLE K C. Gas-phase concentrations of current-use pesticides in Iowa[J]. Environmental Science & Technology, 2005, 39(9): 2952
- [16] KOLPIN D W, NATIONS B K, GOOLSBY D A, et al. Acetochlor in the hydrologic system in the Midwestern United States, 1994[J]. Environmental Science & Technology, 1996, 30(5): 1459
- [17] VOGEL J R, MAJEWSKI M S, CAPEL P D. Pesticides in

- rain in four agricultural watersheds in the United States[J]. Journal of Environmental Quality, 2008, 37(3): 1101
- [18] NEMETH-KONDA L, FÜLEKY G, MOROVJAN G, et al. Sorption behaviour of acetochlor, atrazine, carbendazim, diazinon, imidacloprid and isoproturon on Hungarian agricultural soil[J]. Chemosphere, 2002, 48(5): 545
- [19] HILLER E, CERNANSKÝ S, KRASCSENITS Z, et al. Effect of soil and sediment composition on acetochlor sorption and desorption[J]. Environmental Science and Pollution Research International, 2009, 16(5): 546
- [20] BEDMAR F, DANIEL P E, COSTA J L, et al. Sorption of acetochlor, S-metolachlor, and atrazine in surface and subsurface soil horizons of Argentina[J]. Environmental Toxicology and Chemistry, 2011, 30(9): 1990
- [21] KONDA L N, PÁSZTOR Z. Environmental distribution of acetochlor, atrazine, chlorpyrifos, and propisochlor under field conditions[J]. Journal of Agricultural and Food Chemistry, 2001, 49(8): 3859
- [22] SUN X Y, ZHOU Q X, REN W J, et al. Spatial and temporal distribution of acetochlor in sediments and riparian soils of the Songhua River basin in northeastern China[J]. Journal of Environmental Sciences, 2011, 23(10): 1684
- [23] BAKKE B, DE ROOS A J, BARR D B, et al. Exposure to atrazine and selected non-persistent pesticides among corn farmers during a growing season[J]. Journal of Exposure Science & Environmental Epidemiology, 2009, 19(6): 544
- [24] ARCURY T A, GRZYWACZ J G, BARR D B, et al. Pesticide urinary metabolite levels of children in eastern North Carolina farmworker households[J]. Environmental Health Perspectives, 2007, 115(8): 1254
- [25] ARCURY T A, GRZYWACZ J G, ISOM S, et al. Seasonal variation in the measurement of urinary pesticide metabolites among Latino farmworkers in eastern North Carolina[J]. International Journal of Occupational and Environmental Health, 2009, 15(4): 339
- [26] CHEVRIER C, SERRANO T, LECERF R, et al. Environmental determinants of the urinary concentrations of herbicides during pregnancy: the PELAGIE mother-child cohort (France)[J]. Environment International, 2014, 63:
- [27] CHEVRIER C, LIMON G, MONFORT C, et al. Urinary biomarkers of prenatal atrazine exposure and adverse birth outcomes in the PELAGIE birth cohort[J]. Environmental Health Perspectives, 2011, 119(7): 1034
- [28] WICKERHAM E L, LOZOFF B, SHAO J, et al. Reduced birth weight in relation to pesticide mixtures detected in cord blood of full-term infants[J]. Environment Inter-

- national, 2012, 47: 80
- [29] HELBING C C, OVASKA K, JI L. Evaluation of the effect of acetochlor on thyroid hormone receptor gene expression in the brain and behavior of Rana catesbeiana tadpoles[J]. Aquatic Toxicology, 2006, 80(1): 42
- [30] HINTHER A, DOMANSKI D, VAWDA S, et al. C-fin: a cultured frog tadpole tail fin biopsy approach for detection of thyroid hormone-disrupting chemicals[J]. Environmental Toxicology and Chemistry, 2010, 29(2): 380
- [31] LI W, ZHA J M, LI Z L, et al. Effects of exposure to acetochlor on the expression of thyroid hormone related genes in larval and adult rare minnow (*Gobiocypris rarus*)

 [J]. Aquatic Toxicology, 2009, 94(2): 87
- [32] ROLLEROVÁ E, GÁSPÁROVÁ Z, WSÓLOVÁ L, et al. Interaction of acetochlor with estrogen receptor in the rat uterus. Acetochlor: possible endocrine modulator?[J]. General Physiology and Biophysics, 2000, 19(1): 73
- [33] ROLLEROVA E, WSOLOVA L, URBANCIKOVA M. Neonatal exposure to herbicide acetochlor alters pubertal development in female wistar rats[J]. Toxicology Mechanisms and Methods, 2011, 21(5): 406
- [34] CRUMP D, WERRY K, VELDHOEN N, et al. Exposure to the herbicide acetochlor alters thyroid hormone-dependent gene expression and metamorphosis in *Xenopus Laevis*[J]. Environmental Health Perspectives, 2002, 110(12): 1199
- [35] CHEEK A O, IDE C F, BOLLINGER J E, et al. Alteration of leopard frog (*Rana Pipiens*) metamorphosis by the herbicide acetochlor[J]. Archives of Environmental Contamination and Toxicology, 1999, 37(1): 70
- [36] 赵京津, 杜书明, 张瑶鑫. 人血清中乙草胺的气相色谱测定法[J]. 中华劳动卫生职业病杂志, 2010, 28(8): 620
- [37] WESTER R C, MELENDRES J L, MAIBACH H I. *In Vivo* percutaneous absorption of acetochlor in the rhesus monkey: dose-response and exposure risk assessment[J]. Food and Chemical Toxicology, 1996, 34(10): 979
- [38] DONG H T, XU D D, HU Y, et al. Erythema multiformelike eruption following acute allergic contact dermatitis after exposure to the emulsified herbicide acetochlor[J]. Contact Dermatitis, 2014, 71(3): 178
- [39] CUI Z J, HABARA Y, SATOH Y I. Photodynamic modulation of adrenergic receptors in the isolated rat hepatocytes[J]. Biochemical and Biophysical Research Communications, 2000, 277(3): 705
- [40] CUI Z J, GUO L L. Assessing physiological concentrations of endogenous substances *in situ* by inducing calcium oscillations *in vitro*. Case of liver[J]. Acta Pharmacologica Sinica, 2002, 23(1): 27
- [41] CUI Z J, GUO L L. Photodynamic modulation by Victoria

- Blue BO of phenylephrine-induced calcium oscillations in freshly isolated rat hepatocytes[J]. Photochemical & Photobiological Sciences, 2002, 1(12): 1001
- [42] GASPERS L D, MÉMIN E, THOMAS A P. Calcium-dependent physiologic and pathologic stimulus-metabolic response coupling in hepatocytes[J]. Cell Calcium, 2012, 52(1): 93
- [43] LIANG H Y, SONG Z M, CUI Z J. Lasting inhibition of receptor-mediated calcium oscillations in pancreatic acini by neutrophil respiratory burst: a novel mechanism for secretory blockade in acute pancreatitis?[J]. Biochemical and Biophysical Research Communications, 2013, 437(3): 361
- [44] AN Y P, XIAO R, CUI H, et al. Selective activation by photodynamic action of cholecystokinin receptor in the freshly isolated rat pancreatic acini[J]. British Journal of Pharmacology, 2003, 139(4); 872
- [45] FANG X F, CUI Z J. The anti-botulism triterpenoid toosendanin elicits calcium increase and exocytosis in rat sensory neurons[J]. Cellular and Molecular Neurobiology, 2011, 31(8): 1151
- [46] JIA Y H, CUI Z J. Tri-phasic modulation of ACh- and NA-maintained calcium plateau by high potassium in isolated mouse submandibular granular convoluted tubular cells[J]. Archives of Oral Biology, 2011, 56(11): 1347
- [47] MA C Y, CHEN C Y, CUI Z J. Selective use of a reserved mechanism for inducing calcium oscillations[J]. Cellular Signalling, 2004, 16(12): 1435
- [48] ZHOU Y D, FANG X F, CUI Z J. UVA-induced calcium oscillations in rat mast cells[J]. Cell Calcium, 2009, 45(1):18
- [49] LI Z Y, JIANG W Y, CUI Z J. An essential role of NAD(P)H oxidase 2 in UVA-induced calcium oscillations in mast cells[J]. Photochemical & Photobiological Sciences, 2015, 14(2): 414
- [50] OZCAN L, WONG C C L, LI G, et al. Calcium signaling through CaMKII regulates hepatic glucose production in fasting and obesity[J]. Cell Metabolism, 2012, 15(5): 739
- [51] WANG Y G, LI G, GOODE J, et al. Inositol-1, 4, 5-trisphosphate receptor regulates hepatic gluconeogenesis in fasting and diabetes[J]. Nature, 2012, 485(7396): 128
- [52] LAGOUDAKIS L, GARCIN I, JULIEN B, et al. Cytosolic calcium regulates liver regeneration in the rat[J]. Hepatology, 2010, 52(2): 602
- [53] EVRARD E, MARCHAND J, THERON M, et al. Impacts of mixtures of herbicides on molecular and physiological responses of the European flounder *Platichthys flesus*[J]. Comparative Biochemistry and Physiology Toxicology & Pharmacology: CBP, 2010, 152(3): 321

- [54] ASHBY J, KIER L, WILSON A G, et al. Evaluation of the potential carcinogenicity and genetic toxicity to humans of the herbicide acetochlor[J]. Human & Experimental Toxicology, 1996, 15(9): 702
- [55] ASHBY J, TINWELL H, LEFEVRE P A, et al. Evaluation of the mutagenicity of acetochlor to male rat germ cells[J].

 Mutation Research/Genetic Toxicology and Environmental Mutagenesis, 1997, 393(3): 263
- [56] GREEN T, LEE R, MOORE R B, et al. Acetochlor-induced rat nasal tumors: further studies on the mode of action and relevance to humans[J]. Regulatory Toxicology and Pharmacology: RTP, 2000, 32(1): 127
- [57] HURLEY P M. Mode of carcinogenic action of pesticides inducing thyroid follicular cell tumors in rodents[J]. Environmental Health Perspectives, 1998, 106(8): 437
- [58] AHMED A, MUKHERJEE S, DEOBAGKAR M, et al. Rapid burst of H₂O₂ by plant growth regulators increases intracellular Ca²⁺ amounts and modulates CD4+ T cell activation[J]. International Immunopharmacology, 2010, 10(11): 1397
- [59] ZAFEIRIDOU G, GERONIKAKI A, PAPAEFTHIMIOU C, et al. Assessing the effects of the three herbicides acetochlor, 2, 4, 5-trichlorophenoxyacetic acid (2, 4, 5-T) and 2, 4-dichlorophenoxyacetic acid on the compound action potential of the sciatic nerve of the frog (*Rana ridibunda*)[J]. Chemosphere, 2006, 65(6): 1040
- [60] KOCSIS Z, MARCSEK Z L, JAKAB M G, et al. Chemopreventive properties of trans resveratrol against the cytotoxicity of chloroacetanilide herbicides *in vitro*[J]. International Journal of Hygiene and Environmental Health, 2005, 208(3): 211
- [61] ZHANG J N, LIANG W J, WU X, et al. Toxic effects of acetochlor on mortality, reproduction and growth of *Caenorhabditis elegans* and *Pristionchus pacificus*[J]. Bulletin of Environmental Contamination and Toxicology, 2013, 90(3): 364
- [62] HE H Z, CHEN G K, YU J, et al. Individual and joint toxicity of three chloroacetanilide herbicides to freshwater cladoceran *Daphnia carinata*[J]. Bulletin of Environmental Contamination and Toxicology, 2013, 90(3): 344
- [63] RICH J D, GABRIEL S M, SCHULTZ-NORTON J R. *In vitro* effects of herbicides and insecticides on human breast cells[J]. ISRN Toxicology, 2012, 2012; 232461
- [64] BERNASINSKA J, DUCHNOWICZ P, KOTER-MICHALAK M, et al. Effect of safeners on damage of human erythrocytes treated with chloroacetamide herbicides[J]. Environmental Toxicology and Pharmacology, 2013, 36(2): 368
- [65] ZERIN T, SONG H Y, KIM Y S. Extracellular signal-

- regulated kinase pathway play distinct role in acetochlor-mediated toxicity and intrinsic apoptosis in A549 cells[J]. Toxicology *in Vitro*, 2015, 29(1): 85
- [66] RAYBURN A L, MOODY D D, FREEMAN J L. Cytotoxicity of technical grade versus formulations of atrazine and acetochlor using mammalian cells[J]. Bulletin of Environmental Contamination and Toxicology, 2005, 75(4): 691
- [67] PATEL S, ROBB-GASPERS L D, STELLATO K A, et al. Coordination of calcium signalling by endothelial-derived nitric oxide in the intact liver[J]. Nature Cell Biology, 1999, 1(8): 467
- [68] BARTLETT P J, GASPERS L D, PIEROBON N, et al. Calcium-dependent regulation of glucose homeostasis in the liver[J]. Cell Calcium, 2014, 55(6): 306
- [69] HUANG T, HUANG Y, HUANG Y, et al. Toxicity

- assessment of the herbicide acetochlor in the human liver carcinoma (HepG2) cell line[J]. Chemosphere, 2020, 243: 125345
- [70] WANG S S, ZHANG Y, GAO J F, et al. The enantioselective study of the toxicity effects of chiral acetochlor in HepG2 cells[J]. Ecotoxicology and Environmental Safety, 2021, 218: 112261
- [71] JEFFERIES P R, QUISTAD G B, CASIDA J E. Dialkylquinonimines validated as *in vivo* metabolites of alachlor, acetochlor, and metolachlor herbicides in rats[J]. Chemical Research in Toxicology, 1998, 11(4): 353
- [72] COLEMAN S, LINDERMAN R, HODGSON E, et al. Comparative metabolism of chloroacetamide herbicides and selected metabolites in human and rat liver microsomes[J]. Environmental Health Perspectives, 2000, 108(12); 1151

除草剂乙草胺抑制大鼠肝脏细胞肾上腺素能受体所 介导的胞浆钙振荡

刘方圆1) 张彬彬1,2) 崔宗杰1)

(1)北京师范大学细胞生物学研究所,100875,北京;

2)滨州医学院药学院生物技术实验教学中心, 264003, 山东烟台)

摘要 田间除草剂乙草胺施用之后非常容易进入地表水和地下水,河流溪水中乙草胺浓度可达纳摩尔水平,从而对水生生物产生长期影响. 因职业关系而产生的皮肤暴露和吸入,可导致人血液乙草胺浓度达到微摩尔水平. 对乙草胺体内动力学的研究,发现肝脏是乙草胺毒理作用的主要靶器官. 已知在肝脏细胞多种生理功能中,钙离子发挥重要作用. 本文研究在新鲜分离的大鼠肝脏细胞,乙草胺对肾上腺素能受体所介导胞浆钙振荡的可能影响. 实验发现低浓度乙草胺 $(1,10\ \mu mol\cdot L^{-1})$ 对苯丙肾上腺素所诱导钙振荡没有影响,但是高浓度 $(50,100,200\ \mu mol\cdot L^{-1})$ 乙草胺在有些肝脏细胞可逆性抑制苯丙肾上腺素所诱导的胞浆钙振荡. 在苯丙肾上腺素 2 次串联刺激之间短暂加入乙草胺 $(1,10,100\ \mu mol\cdot L^{-1})$,乙草胺对肝脏细胞基础钙浓度没有影响,也不影响第 2 次苯丙肾上腺素刺激所引发胞浆钙振荡. 细胞免疫化学研究发现在新鲜分离的大鼠肝脏不同细胞, α 1 肾上腺素能受体密度存在明显差异. 对固定的大鼠肝脏切片进行组织免疫化学检测,发现 α 1 肾上腺素能受体在肝脏小叶的密度梯度分布: α 1 肾上腺素能受体密度从中央静脉方向到门脉方向从低到高逐渐增加. 这种 α 1 肾上腺素能受体的密度梯度分布,可能是乙草胺选择性抑制苯丙肾上腺素刺激分离大鼠肝脏细胞所诱发钙振荡的原因. 本文数据揭示,在急性皮肤暴露或乙草胺中毒后可达到的血液浓度,短期施加乙草胺,可一过性抑制肝脏 α 1 肾上腺素能受体所介导的胞浆钙振荡. 长期乙草胺暴露,可能对肝脏细胞正常钙信号系统产生持续性不良影响.

关键词 乙草胺;肾上腺素能受体;钙振荡;肝脏细胞中图分类号 Q2; S482

DOI: 10.12202/j.0476-0301.2022145

【责任编辑:武 佳】